

Rapid temporal adaptation structures tolerance to toxic cyanobacteria in a natural population of the water flea *Daphnia*

Maxime Fajgenblat^{1,2}, Emma Gouwy³, Manon Coone³, Rafaela Almeida¹, Alice Boudry^{1,3}, Kiani Cuypers¹, Edwin van den Berg¹, Isabel Vanoverberghe³, Luc De Meester^{1,4,5,6}, Ellen Decaestecker³

¹Laboratory of Freshwater Ecology, Evolution and Conservation, KU Leuven, Leuven, Belgium

²Data Science Institute, I-BioStat, Hasselt University, Diepenbeek, Belgium

³Laboratory of Aquatic Biology, Interdisciplinary Research Facility Life Sciences, KU Leuven, KULAK, Campus Kortrijk, Kortrijk, Belgium

⁴Leibniz Institute für Gewässerökologie und Binnenfischerei (IGB), Berlin, Germany

⁵Institute of Biology, Freie Universität Berlin, Berlin, Germany

⁶Berlin-Brandenburg Institute of Advanced Biodiversity Research (BBIB), Berlin, Germany

Corresponding author: Laboratory of Freshwater Ecology, Evolution and Conservation, KU Leuven, BE-3000, Leuven, Belgium. Email: maxime.fajgenblat@gmail.com
M.F. and E.G. contributed equally.

L.D.M. and E.D. contributed equally.

Abstract

Cyanobacteria blooms pose a substantial threat to freshwater systems globally. While zooplankton grazers such as *Daphnia* can have an important role in suppressing cyanobacteria blooms, cyanobacteria can adversely impact *Daphnia* fitness and even kill them. Earlier work has shown an evolutionary increase in tolerance to cyanobacteria across years and strong genotype $\times$ genotype interactions determining the interaction between *Daphnia* and the cyanobacterium *Microcystis*. Here, we test the hypothesis that *Daphnia magna* can adapt during 1 growing season to changes in dominant strains of *Microcystis*. Over 2 consecutive years, we collected *D. magna* clonal lineages and *Microcystis* strains from a single pond early and late in the growing season and we assessed whether *Daphnia* survival differed when exposed to *Microcystis* strains from either the same or a different time point within the growth season. Our findings reveal important *Daphnia* genotype $\times$ *Microcystis* genotype interactions, with *Daphnia* survival being higher when exposed to *Microcystis* from the same time point than when exposed to *Microcystis* of a different time point. Our results extend earlier findings to variation within 1 single natural system and growth season, and suggest an important impact of rapid (co)evolutionary dynamics shaping the tolerance of zooplankton grazers to cyanobacteria.

Keywords: coevolution, eco-evolutionary dynamics, genotype $\times$ genotype interactions, harmful algal blooms, *Daphnia*, top-down control

Introduction

Cyanobacteria blooms are increasingly common across the globe because of rising temperatures and nutrient levels (Kosten et al., 2012; Meerhoff et al., 2022). Cyanobacteria blooms can exert substantial pressure on ecosystem functioning due to their effect on the fitness of many organisms (Pimentel et al., 2001). They reduce nutrient and light availability for submerged macrophytes, can typically not be grazed because they form colonies, are generally a low-quality food source for grazers, and can produce toxins (Carton, 2005; Gray et al., 2014; Paerl et al., 2011; Sommer et al., 2012). As such, they can shift shallow lakes to a turbid state dominated by cyanobacteria, which is difficult to control by grazers (Hansson et al., 2007; Scheffer & Van Nes, 2007). In temperate regions, cyanobacteria blooms typically occur in summer (PEG model; Sommer & Maciej Gliwicz, 1986). Higher temperatures, along with nutrient release from the sediment (Paerl & Huisman,

2008) and decreased zooplankton grazing caused by fish predation (Lampert & Sommer, 2007), lead to an increased fraction of cyanobacteria in the phytoplankton community.

The water flea genus *Daphnia* plays a pivotal role in many standing freshwaters, as they are key grazers on phytoplankton and the preferred prey of fish (Miner et al., 2012). *Daphnia* can often exert a strong top-down control on phytoplankton (Gianuca et al., 2016; Jeppesen et al., 1996), and can even make the difference between a cyanobacteria bloom and a clear-water state (Sarnelle & Wilson, 2005; Scheffer et al., 2001). Yet, many studies have reported that the cyanobacterium *Microcystis*, a cosmopolitan genus known for its bloom-forming behavior and significant ecological and human health impacts (Paerl & Otten, 2013), can kill *Daphnia* due to the production of toxins or feeding inhibition and starvation, for instance, through its low nutritional quality, the production of digestive protease inhibitors, and colony forma-

Received February 7, 2025; revisions revised May 26, 2025; accepted May 30, 2025

© The Author(s) 2025. Published by Oxford University Press on behalf of The Society for the Study of Evolution (SSE) and European Society for Evolutionary Biology (ESEB).

This is an Open Access article distributed under the terms of the Creative Commons Attribution-NonCommercial-NoDerivs licence (<https://creativecommons.org/licenses/by-nc-nd/4.0/>), which permits non-commercial reproduction and distribution of the work, in any medium, provided the original work is not altered or transformed in any way, and that the work is properly cited. For commercial re-use, please contact reprints@oup.com for reprints and translation rights for reprints. All other permissions can be obtained through our RightsLink service via the Permissions link on the article page on our site-for further information please contact journals.permissions@oup.com

tion (Agrawal et al., 2001; Ghadouani et al., 2004; Lüring, 2003; Macke et al., 2017; Trubetskova & Haney, 2006; von Elert et al., 2003). Consequently, while *Daphnia* can serve as an efficient grazer of *Microcystis*, it can also be vulnerable to it. This leads to a large variation in results on top-down control of cyanobacteria blooms by *Daphnia*, as reported in field-based and experimental work (e.g., Benndorf & Henning, 1989; de Bernardi & Giussani, 1990; Dejenie et al., 2009; Gulati et al., 2008; Sarnelle & Wilson, 2005; Søndergaard et al., 2008). Several studies indicate that *Daphnia* displays inducible responses (which can be transferred through maternal effects; Gustafsson et al., 2005) and can even genetically adapt when exposed to cyanobacteria (e.g., through selection on digestive proteases; Schwarzenberger et al., 2020), which may contribute to the variation in empirical outcomes. Lemaire et al. (2011) experimentally showed both variation in tolerance to *Microcystis* among *Daphnia* genotypes and variation in overall toxicity of *Microcystis* strains, but also striking genotype $\times$ genotype interactions that make the observed survival of *Daphnia* when exposed to *Microcystis* dependent on the specific combination of *Daphnia* and *Microcystis* genotypes. Such genotype $\times$ genotype interactions are often observed in host-parasite interactions and can lead to rapid local adaptation and Red Queen dynamics (Capaul & Ebert, 2003; Decaestecker et al., 2007; Lambrechts et al., 2006). A resurrection study has indeed shown that *Daphnia* in Lake Constance evolved higher tolerance to cyanobacteria as the lake eutrophied (Hairston et al., 1999, 2001) and subsequently lost this adaptation as nutrient concentrations in the lake became lower again (Isanta-Navarro et al., 2021). In line with these results, Sarnelle & Wilson (2005) observed that *Daphnia pulicaria* populations from eutrophied ponds are significantly more tolerant to cyanobacteria than *D. pulicaria* populations from less eutrophic ponds.

Because of their role as key grazers of phytoplankton, the capacity of *Daphnia* populations to evolve enhanced tolerance to cyanobacteria can have important ecological consequences. By combining field monitoring with genotyping and laboratory experiments, Schaffner et al. (2019) showed that *Daphnia mendotae* clones dominant in summer are better able to feed on low-quality algae communities with a high proportion of cyanobacteria and that this evolutionary change increased their abundance and thus the potential for top-down control of phytoplankton. Hegg et al. (2022) did not find evidence for increased tolerance to cyanobacteria when comparing clones obtained before and after a bloom. In this experiment of Hegg et al. (2022), however, standard laboratory strains were used rather than local strains of *Microcystis*. If genotype $\times$ genotype interactions are important and lead to patterns of local (co)adaptation, the use of laboratory strains of *Microcystis* might fail to reveal the resulting eco-(co)evolutionary dynamics.

In this study, we set out to test whether adaptation of *Daphnia* populations to changes in *Microcystis* strain composition within a single growing season can be an important aspect of eco-(co)evolutionary dynamics that determine the outcome of the interaction between *Daphnia* and *Microcystis*. More specifically, we tested whether a *Daphnia magna* population inhabiting a shallow lake can genetically track changes in *Microcystis* strain composition through rapid evolution as a bloom develops in spring. Over two consecutive years, we harvested *Daphnia* and *Microcystis* from two time points in the growing season. We subsequently performed a reciprocal exposure experiment where *Daphnia* clonal lineages were exposed to *Microcystis* strains from either the same or the other time point. We tested the hypotheses that (i) *Daphnia* genotype $\times$ *Microcystis* genotype interactions occur with the capacity of *Daphnia* to cope with *Microcystis* within the set of genotypes and strains isolated from a single lake, and (ii) *Daphnia* clonal

lineages are more tolerant to *Microcystis* strains isolated during the same moment in the growing season than to *Microcystis* strains isolated from a different moment in the growing season.

Methods

Isolation of *Daphnia* and *Microcystis*

We isolated 12 clonal lineages of *D. magna* and 12 isolates of *Microcystis* sp. from Langerodevijver, a man-made pond located in a nature reserve in Belgium (50°49'42.9" N; 4°38'23.8" E). To capture the interannual and seasonal dynamics within Langerodevijver, isolation took place in two consecutive years (2018 and 2019) during two distinct time periods that we will further refer to as "early" and "late" in the growing season: end of April 2018, end of May 2018, end of April 2019, and mid to end of June 2019 (Table S1). These time points were chosen because of the temporal dynamics of phytoplankton, cyanobacteria, and different *Daphnia* species in this pond (see Figure S1), where *Daphnia* appears in samples around mid-March and disappears from the pond around June, presumably because of predation by young-of-the-year fish. Cyanobacteria are still at low densities, and generally start to bloom after the disappearance of *D. magna*; as a result all isolations of *Daphnia* and *Microcystis* were carried out before the peak annual *Microcystis* sp. summer bloom. At each of the four sampling moments, three *D. magna* females were isolated from the water column to establish clonal lineages and a water sample was taken and stored at 4 °C for isolation of *Microcystis* sp. strains.

Daphnia magna cultures

Daphnia magna individuals were individually placed in 150 ml glass jars filled with filtered tap water (Greenline e1902 filter) to establish clonal lineages, which were maintained under standardized laboratory conditions prior to the experiment (16:8 hr light-dark cycle, constant ambient temperature of 19 ± 1 °C, fed three times a week 200×10^3 cells/ml of *Chlorella vulgaris*, and medium refreshed once a month). These conditions ensure the maintenance of clonal lines, without interference of recombination due to sexual reproduction. Before the start of the experiment, all clonal lines were kept under optimal conditions for at least two generations to exclude interference of maternal effects and to ensure that the observed responses were due to genetic effects. Animals were kept individually in 50 ml of aged tap water (16:8 hr light-dark cycle, constant ambient temperature of 19 ± 1 °C, fed three times a week 200×10^3 cells/ml of *Chlorella vulgaris*, and medium refreshed once a week). Both the maternal generation and the generation of experimental animals were initiated by inoculating one neonate <24 hr old from the second or third brood.

Microcystis sp. cultures

For the isolation of *Microcystis* sp., three droplets of sampled water were placed on a glass microscope slide and placed under an SZX10 stereomicroscope (Olympus, Tokyo, Japan). Single colonies were then picked out using the following mouth-pipetting technique: the tip of a glass Pasteur pipette was heated using a Bunsen flame and then extended using forceps. This extended pipette was then connected to an autoclaved flexible, plastic tube. The other end of the tube was placed in the mouth of the person isolating the cells. Single colonies were identified under the microscope and carefully pipetted up by exerting gentle suction on the plastic tube end. The isolated colony was then washed three times in modified Wright's Cryptophyte medium (i.e., without Tris and a fivefold of NaNO_3 and K_2HPO_4 , supplemented with vitamins), known to pro-

mote *Microcystis* sp. growth (Guillard & Lorenzen, 1972). The single *Microcystis* sp. colony was then placed in 2 ml of this medium in a 24-well plate. Plates were covered with parafilm to avoid contamination and placed on a shaker under a 16:8 hr light cycle at 21 ± 1 °C. Additionally, plates were covered with a neutral density light filter to reduce light intensity and to stimulate *Microcystis* sp. growth. Further manipulation of the cyanobacterial cultures was performed under the flow hood. Once the 2 ml culture showed a healthy green color, it was added to 5 ml of fresh medium in a six-well plate under the flow hood. The empty well was rinsed thoroughly, and the new plate was covered with parafilm and placed on a shaker. Once the cultures were dense, three cultures per time point were selected and individually transferred to a 1 L glass culture bottle (Düran) under aeration, with the bottle caps fitted with 0.22 μ m Millipore air filters. At the aeration system's input and output to supply CO₂ and to remove oxygen, a second 0.22 μ m filter was used to prevent (cyano)bacterial contamination. To maintain regular mixing and prevent algal precipitation, magnetic stirrers were used. The algal culture cell density was assessed using flow cytometry on the FACS Verse cell sorter (Biosciences). Prior to cell counting, colonies were separated into individual cells by vigorously shaking the vials using a vortex mixer.

Experimental setup

The experiment was conducted in February 2021. *Daphnia magna* clonal lineages were administered *Microcystis* sp. strains from either the same (further referred to as “contemporal”) or a different time point (further referred to as “allotemporal”) within the same year (Figure S2). We did this for *D. magna* clonal lineages and *Microcystis* sp. strains from Langerodevijver isolated in two consecutive years, enabling us to assess the repeatability of the observed pattern over time. We considered 72 combinations of *Daphnia* clonal lineages and *Microcystis* strains (2 years $\times$ 2 *Daphnia* populations $\times$ 3 *Daphnia* clonal lineages per population exposed to 2 *Microcystis* populations $\times$ 3 *Microcystis* strains per population) as experimental units, each including six experimental animals. The contemporal and allotemporal combinations of one *Daphnia* clonal lineage (4b) were excluded because it did not produce sufficient offspring, leading to a reduction by 6 units. Two other clonal lineages (1c and 4a) did not produce sufficient brood to use six experimental animals, leading to 11 experimental units that only contained four or five individuals. Consequently, 66 out of the 72 experimental units with a total of 380 experimental *Daphnia* individuals were effectively executed. The *Daphnia* individuals were cultured for a total of 12 days, since this duration has been shown to highlight clonal differences in sensitivity to cyanobacteria (Macke et al., 2017). Starting from day 1 of the experiment, 30×10^3 cells/ml of the corresponding *Microcystis* sp. strain were administered every 4 days. Survival was monitored on days 5, 9, and 12 of the experiment (Figure S3). Experimental units were kept at 19 ± 1 °C, 16:8 hr light–dark regime.

Statistical analyses

Throughout the following, we refer to “*Daphnia* timing” as the variable indicating whether early (0) or late (1) *Daphnia* are considered, to “*Microcystis* timing” as the variable indicating whether *Daphnia* are exposed to early (0) or late (1) *Microcystis* strains and to “contemporality” as the variable indicating whether *Daphnia* are exposed to *Microcystis* with a different (0, allotemporal) or matching timing (1, contemporal).

As our main analysis, we use a binomial generalized linear mixed model (GLMM) with a logistic link function to model sur-

vival at the 12th day, i.e., the final experimental day. The number of surviving individuals y_i out of the total number of individuals n_i in experimental unit i is modeled as follows:

$$y_i \sim \text{Binomial}(n_i, p_i),$$

$$\text{logit}(p_i) = \beta_0^{(i)} + \text{Daphnia_timing} \cdot \beta_1^{(i)} + \text{Microcystis_timing} \cdot \beta_2^{(i)} + \text{contemporality} \cdot \beta_3^{(i)} + b_{\text{clone}(i)}^{\text{Daphnia}} + b_{\text{strain}(i)}^{\text{Microcystis}},$$

where β_0^j is the intercept, β_1^j is the effect of *Daphnia* timing, β_2^j is the effect of *Microcystis* timing, β_3^j is the effect of contemporality, $b_{\text{clone}(i)}^{\text{Daphnia}}$ is a normally distributed random effect for the *Daphnia* clonal lineage clone(i) used in experimental unit i , and $b_{\text{strain}(i)}^{\text{Microcystis}}$ is a normally distributed random effect of the *Microcystis* strain(i) used in experimental unit i . To allow model parameters to vary across the 2 years, two distinct sets of regression coefficients are used (i.e., one set for 2018 and one set for 2019), with the superscript $j(i)$ referring to the year of experimental unit i . In an alternative model specification, a third random effect b_i^{unit} is included at the level of the experimental unit, to account for any overdispersion that might be caused by, for instance, additional genotype $\times$ genotype interactions unrelated to contemporality left unexplained by additive genotype effects captured by the *Daphnia* clonal lineage and *Microcystis* strain random effects alone.

We implemented this model using the probabilistic programming language Stan (Carpenter et al., 2017) through the “cmdstanr” package in R v.4.3.1 (R Core Team, 2023). Stan performs Bayesian inference by means of a dynamic Hamiltonian Monte Carlo algorithm, a gradient-based Markov chain Monte Carlo sampler (Carpenter et al., 2017). We used zero-centered, weakly informative normal priors with scale 3 for the variables β_0 , β_1 , β_2 , and β_3 . Similarly, we used weakly informative half-normal priors with scale 3 for the scales of the random clonal lineage and strain effects $b_{\text{clone}}^{\text{Daphnia}}$ and $b_{\text{strain}}^{\text{Microcystis}}$. We ran eight chains with 2,000 iterations each, of which the first 1,000 were discarded as warm-up. Model convergence was confirmed both visually by means of traceplots and numerically by means of effective sample sizes, the absence of divergent transitions, and the potential scale reduction factor, for which all parameters had $\hat{R} < 1.01$ (Vehtari et al., 2021). A satisfactory goodness-of-fit was confirmed through posterior predictive checks (Gelman et al., 1996). We performed prior sensitivity analyses to confirm that model results are robust against alternative prior specifications. Traceplots and visualizations of the posterior predictive checks and prior sensitivity analyses are provided in the Supplementary Information (Figures S4–S9). We used the “tidybayes” v.2.3.1 package (Kay, 2022) to visualize the posterior distributions of the inferred effects and their corresponding derived quantities.

To conform readers who are more familiar with frequentist statistics, we also fitted the binomial GLMM using maximum likelihood with the “glmer” function of the “lme4” package v.1.1 (Bates et al., 2015) in R v.4.3.1 (R Core Team, 2023). Rather than estimating two sets of parameters for both years within the same model (which is not readily possible using the considered software) or including interactions between all regression parameters and the year, we estimate two separate models in a year-by-year fashion.

To ensure that our results are insensitive to the choice of survival at day 12 as endpoint, we also performed a Bayesian interval-censored frailty survival analysis (Bogaerts et al., 2017) to model individual survival over the entire experiment's duration. We outline methodological details for this analysis in the Supplementary Information (Section S1).

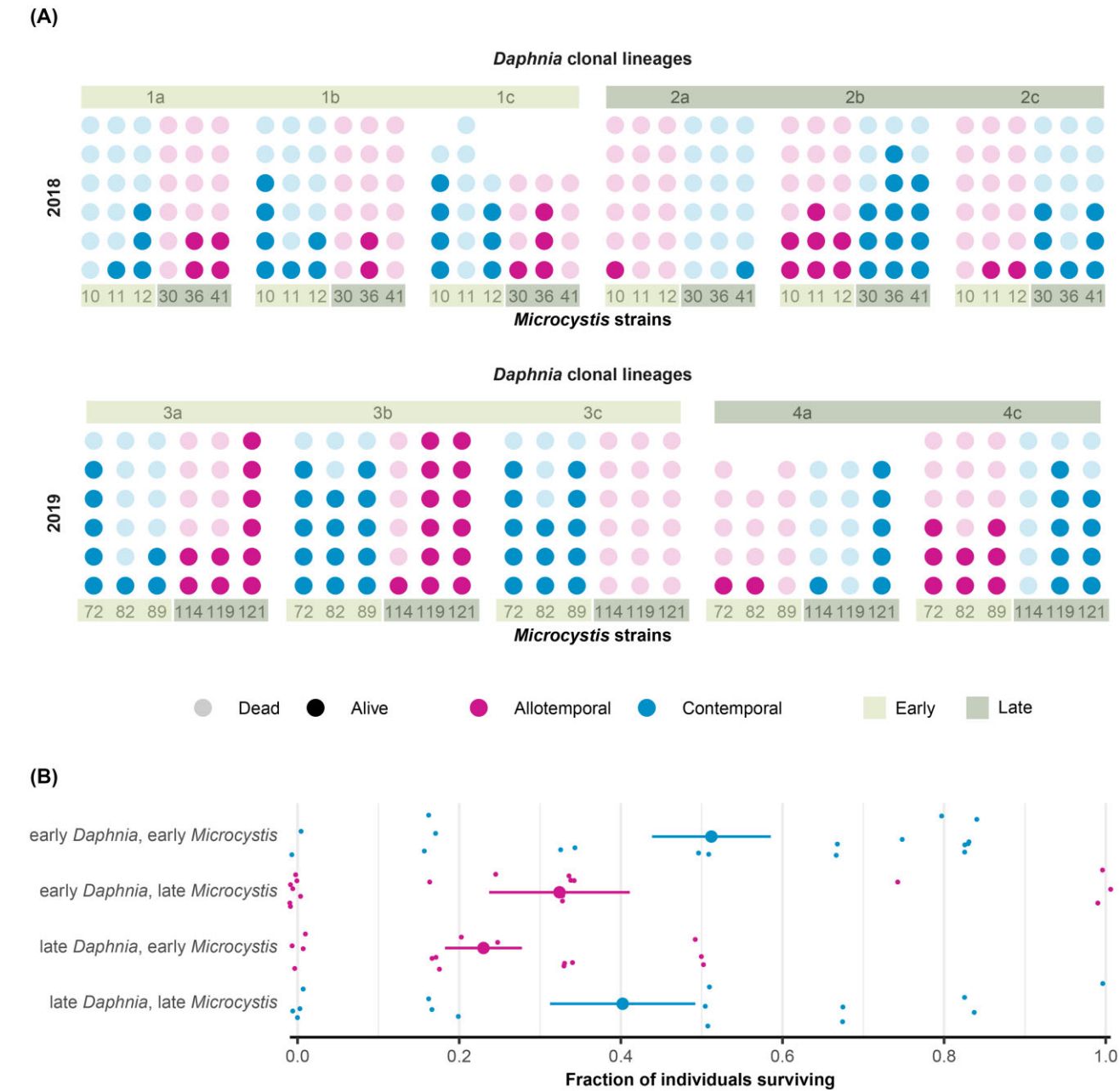


Figure 1. Observed *Daphnia* survival at day 12. (A) Survival status of each individual, represented by a point (dead: transparent; alive: opaque), organized per *Daphnia* clonal lineage, and per *Microcystis* strain. (B) Averaged survival for each combination of *Daphnia* clonal lineage and *Microcystis* strain, jittered horizontally and vertically to a small extent to avoid overlap. Grand means and standard errors for each combination of early and late *Daphnia* and *Microcystis* clonal lineage are shown by means of big dots and horizontal lines.

Results

Daphnia survival after 12 days of exposure to *Microcystis* varied widely across experimental units (Figure 1 and Figure S3). The average observed survival rate across individuals was 37.2%.

Using a binomial GLMM to analyze survival at day 12, we found no evidence for differential survival between early and late *Daphnia* clonal lineages (posterior probability of a positive difference, averaged across both years: 70.9%), neither did we find evidence for a difference between early and late *Microcystis* strains (av. post. prob.: 52.4%). We did, however, find very strong statistical support for a positive effect of contemporality (av. post. prob.: 99.6%), with contemporaneous combinations featuring a higher survival compared to allotemporal ones (Figures 1 and 2). Averaged across both

years, the posterior mean contemporality effect equals 0.95 (95% CrI [0.27,1.67]) on the logit scale. This corresponds to a 17.7% decrease in posterior mean survival probability when *Daphnia* are exposed to *Microcystis* strains from another moment (Figure S10). Moreover, these patterns are strongly replicated across both years, with only 54.7% posterior probability of a positive or negative difference in the effect of contemporality between both years. For reference, a year-by-year frequentist binomial GLMM analysis yields highly similar results, with p-values of 0.006 and 0.013 for the effect of contemporality in 2018 and 2019, respectively (Table S2).

The interval-censored frailty survival analysis, in which data of all time points are integrated, yields qualitatively identical

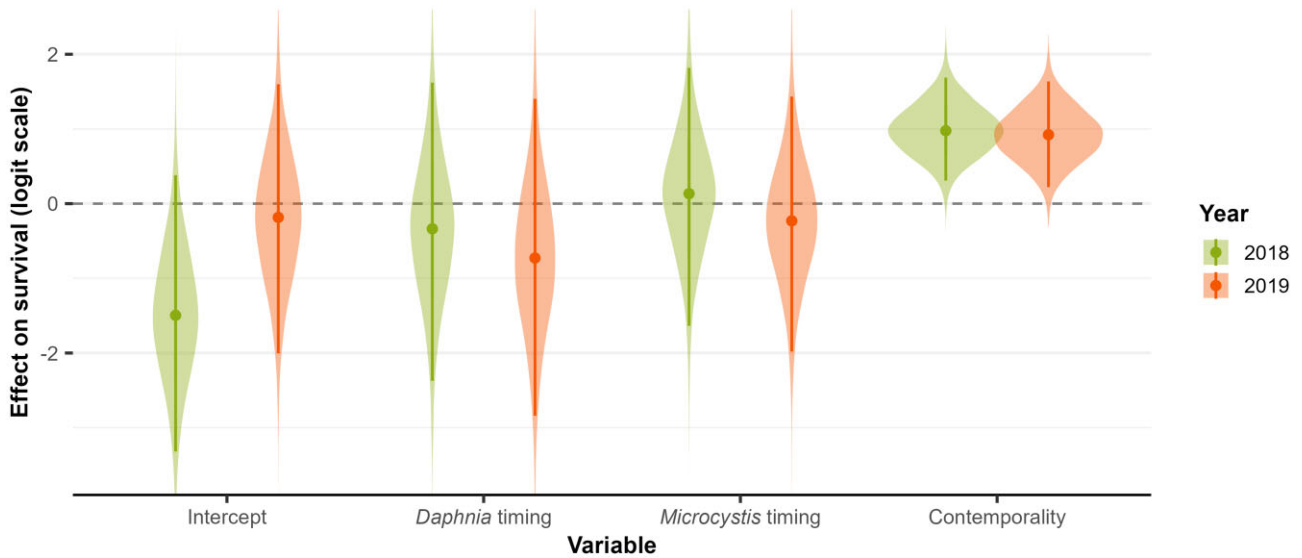


Figure 2. Posterior distributions of the model intercept and the effects of *Daphnia* timing, *Microcystis* timing, and contemporality on *Daphnia* survival (logit scale) at the 12th day, for 2018 and 2019, as obtained from the binomial GLMM. The posterior means and 95% credible intervals are represented by dots and vertical lines, respectively.

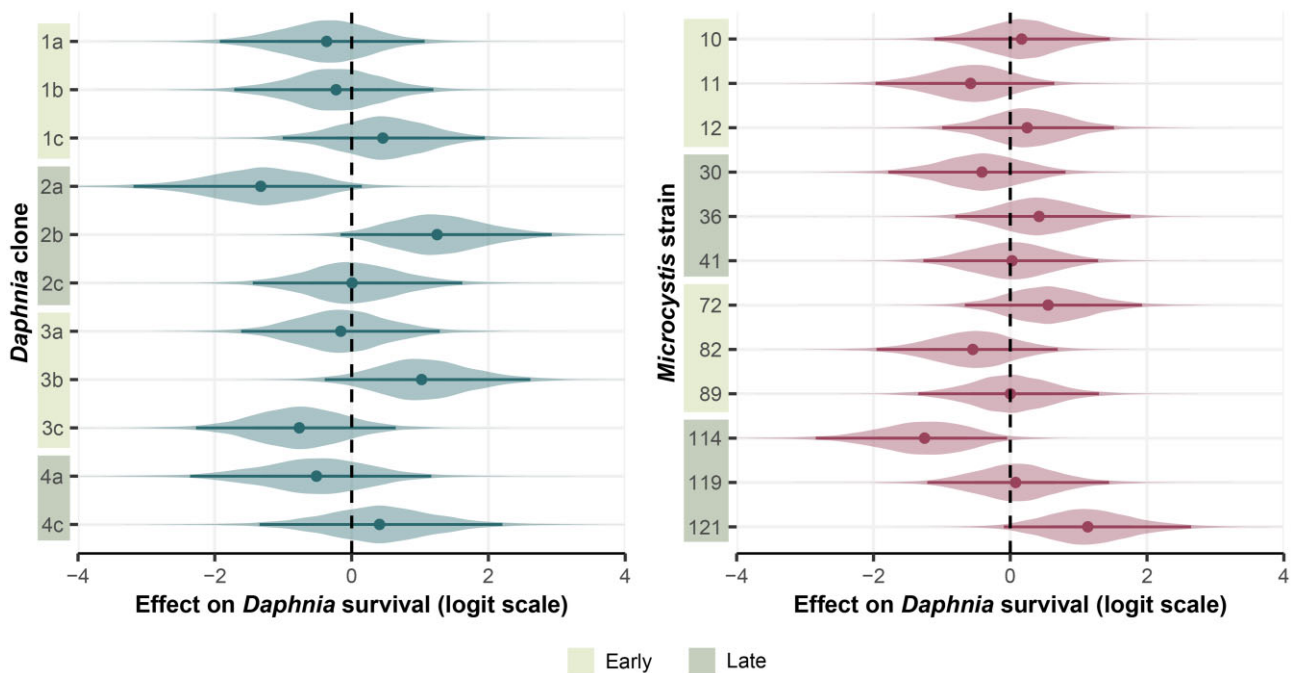


Figure 3. Posterior distributions of the *Daphnia* clonal lineage (left) and *Microcystis* strain (right) effects on *Daphnia* survival (logit scale) at the 12th day, as obtained from the binomial GLMM. These effects represent the clone- and strain-specific variation, corrected for the effects of year, *Daphnia* timing, *Microcystis* timing, contemporality, *Microcystis* strain (for the *Daphnia* clonal lineage effects only), and *Daphnia* clone (for the *Microcystis* strain effects only). The posterior means and 95% credible intervals are represented by dots and horizontal lines, respectively.

results as the binomial GLMM analysis performed on the 12th day (Figure S11). The survival functions generated through this interval-censored frailty survival analysis visually confirm that contemporal combinations host an increased survival compared to allotemporal combinations, while there is much more overlap in confidence intervals among survival curves of early and late *Daphnia* or of *Daphnia* exposed to early and late *Microcystis* (Figure S12).

Posterior distributions for the different *Daphnia* genotypes and *Microcystis* strains show that there is much variation in survival

upon exposure to *Microcystis* among *Daphnia* genotypes (Figures 3 and 4). Similarly, there is variation in survival of *Daphnia* when exposed to different *Microcystis* strains, reflecting that *Microcystis* genotypes differ in their adverse effects toward *Daphnia* (Figures 3 and 4). For instance, *Daphnia* clonal lineages 2a and 2b feature the highest intrinsic difference in tolerance (Figures 3 and 4; posterior probability of a difference: >99.9%). On the other hand, *Daphnia* individuals exposed to *Microcystis* strain 114 have a distinctly lower survival probability at day 12 than individuals exposed to *Microcystis* strain 121 (posterior probability of a difference: >99.9%).

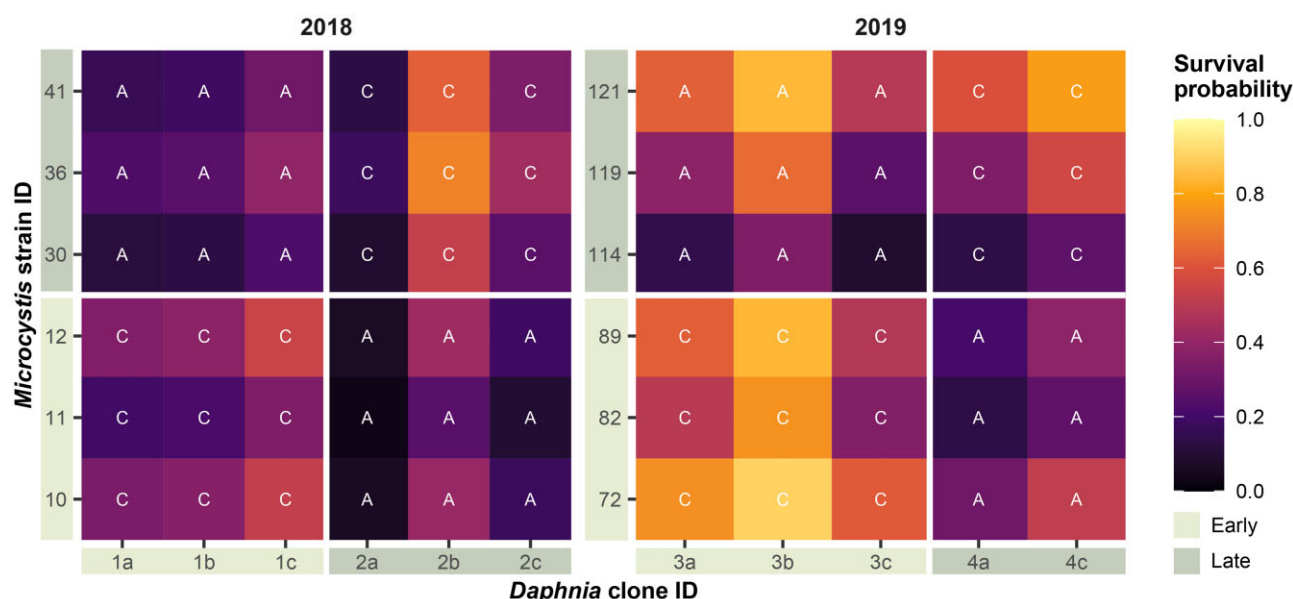


Figure 4. Heatmap showing the posterior mean modeled survival probability for each combination of *Daphnia* clonal lineages and *Microcystis* strains, as obtained from the binomial GLMM. The white letters indicate whether the combination of *Daphnia* clonal lineage and *Microcystis* strain pertains to a contemporal ("C") or allotemporal ("A") condition.

A comparison of the main binomial GLMM with a fully saturated binomial GLMM, in which an observation-level random effect (i.e., a *Daphnia* genotype $\times$ *Microcystis* genotype random effect) is added, indicates that our main model can explain a posterior mean fraction of 67.3% (95% CrI [53.4,79.2]) of the saturated linear predictor's variation (Figure S13). Hence, approximately two-thirds of the variation in survival can be explained by contemporality, *Daphnia* genotype, and *Microcystis* genotype alone. Approximately one-third of the variation is left unexplained, suggesting that genotype $\times$ genotype effects unrelated to contemporality, or other sources of overdispersion, were also important in determining *Daphnia* survival in our experiment. While our experimental design allows to disentangle additive effects of *Daphnia* genotype and *Microcystis* genotype as well as the genotype $\times$ genotype interaction induced by contemporality, variation induced by additional genotype $\times$ genotype interactions might be confounded with observational noise because of the absence of replicates at the level of *Daphnia* genotype $\times$ *Microcystis* genotype combination.

Discussion

In this study, we exposed early- and late-season genotypes isolated from a natural *D. magna* population over 2 years to *Microcystis* strains isolated from the same two time periods of their respective year of isolation. Our results do not only show that *Daphnia* genotypes differ in sensitivity to *Microcystis* and *Microcystis* strains differ in their adverse effects toward *Daphnia*, but also reveal a striking pattern of higher survival when *Daphnia* genotypes are exposed to *Microcystis* strains that were isolated during the same time period compared to *Daphnia* that were exposed to *Microcystis* strains that were isolated during a different period of the growing season.

Daphnia populations have been shown to be genetically differentiated in their capacity to cope with *Microcystis* (Chislock et al., 2019; Sarnelle & Wilson, 2005; Schwarzenberger et al., 2020) and that *Daphnia* can become more tolerant to *Microcystis* through acclimatization (Chislock et al., 2019; Sarnelle & Wilson, 2005), epigenetic effects (Asselman et al., 2017; Ger et al., 2016), or through

microbiome-mediated acclimatization (Macke et al., 2017). Influential studies have shown that *Daphnia* can genetically adapt to the presence of *Microcystis* over the course of decades in resurrection ecology studies (Hairston et al., 1999; Isanta-Navarro et al., 2021; Jiang et al., 2016). Our results suggest that genetic adaptation of *Daphnia* to *Microcystis* can occur over the course of one growing season and enables *Daphnia* to effectively cope with contemporal *Microcystis* populations. While adapting to changes in the *Microcystis* population, *Daphnia* lose some of their tolerance to past populations of *Microcystis*. This is reflected by the fact that late-season *Daphnia* populations suffer from exposure to the early-season *Microcystis* strains. This rapid adaptation is expected to influence seasonal phytoplankton and *Microcystis* dynamics and might explain the genetic adaptation affecting *D. mendotae* seasonality and phytoplankton biomass in Oneida Lake, as observed by Schaffner et al. (2019). The PEG model explaining seasonal changes in phytoplankton and zooplankton biomass in lakes might in part be driven by the eco-evolutionary interactions revealed for a deep lake by Schaffner et al. (2019) and the shallow pond in our study.

We show strong *Daphnia* genotype $\times$ *Microcystis* genotype interactions. These $G \times G$ interactions are a prerequisite for the pattern of contemporal adaptation that we observe, as it allows the population of the one interacting species to not only adapt to *Microcystis* as a stressor, but to adapt to the specific genotype composition of *Microcystis*. *Microcystis* strains have been shown to strongly differ in their toxicity, both in terms of the capacity to produce microcystins and the amount they produce, but also in terms of the synthesis of other polypeptides with potentially toxic effects (Xu et al., 2023) or secondary metabolites such as digestive protease inhibitors (Schwarzenberger et al., 2020). In addition to variation in toxicity, some *Microcystis* strains also exhibit inducible colony formation in response to grazing, impeding ingestion. Considering this wide variety of toxicity profiles and defense strategies, our results suggest that *Daphnia* genotypes are differentially able to adapt to different *Microcystis* strains, enabling clonal selection. Lemaire et al. (2012) also showed important genotype $\times$ genotype interactions in determining the performance of *Daphnia* when ex-

posed to *Microcystis*. They used *Daphnia* clonal lineages and *Microcystis* strains that were isolated from habitats that were geographically isolated, including *Daphnia* from both the European and African continent. This study shows that the same pattern exists within a single habitat, and contributes to our understanding of seasonality of zooplankton–phytoplankton interactions. Our results also align with the seasonal dynamics in the clonal composition of *D. magna* observed in the field by Schwarzenberger et al. (2013), which might be driven by corresponding dynamics in the cyanobacterial composition and the protease inhibitors they produce (Schwarzenberger et al., 2013). The results of our experiments demonstrate that *Daphnia* is indeed able to seasonally track changes in the composition of cyanobacterial strains.

Our results testify to the large intrapopulation genetic variation that is regularly observed in *Daphnia* (Chaturvedi et al., 2021; De Coninck et al., 2013a; Ilić et al., 2021) and that can explain why local adaptation has been reported in *Daphnia* populations in response to a wide range of environmental gradients (Brans & De Meester, 2018; Geerts et al., 2015), pollution (Almeida et al., 2023; Cuenca-Cambronero et al., 2021; De Coninck et al., 2013b), UV light (Miner & Kerr 2010), including interactions with antagonists such as predation (Stoks et al., 2015), parasites (Decaestecker et al., 2007), and competitors (Steiner, 2004). *Daphnia*–*Microcystis* interactions are especially interesting because both interactors can kill each other if they are sufficiently adapted to their enemy. This may result in fast and reciprocal responses, similar to those reported for host–parasite interactions (Decaestecker et al., 2007, 2013; Gandon et al., 2008). Similar to the strong genotype $\times$ genotype interactions that are often observed (Capaul & Ebert, 2003; Ebert, 2008), we also observe strong *Daphnia* genotype $\times$ *Microcystis* genotype interactions, both at larger spatial scales (Lemaire et al., 2011) as well as within single habitats (this paper).

The strong *Daphnia* genotype $\times$ *Microcystis* genotype interactions might not only enable *Daphnia* to adapt to *Microcystis* but may also allow *Microcystis* populations to adapt to the changes in the *Daphnia* population, leading to continuous dynamics of coevolution (Anderson & May, 1982; Ehrlich & Raven, 1964; Ley et al., 2008). Our results show that *Daphnia* populations seem to be able to track changes in *Microcystis* populations during a growing season (i.e., late-season *Daphnia* genotypes are better adapted to late-season *Microcystis* than early-season *Daphnia* genotypes) and show that this adaptation has a cost in terms of susceptibility to early-season *Microcystis* (i.e., late-season *Daphnia* genotypes are less well adapted to early-season *Microcystis* than early-season *Daphnia* genotypes). The fact that early-season *Daphnia* are more susceptible to late-season *Microcystis* might reflect adaptation of the *Microcystis* population to *Daphnia*, driven by selection against grazing. Given that we did not carry out genomic analyses on the different *Microcystis* strains used in this experiment, we cannot be certain that evolution is involved rather than epigenetic changes or a shift in relative abundances of closely related *Microcystis* species. As an alternative to the interpretation of the observed shifts being caused by selection on standing genetic variation present in the hatched population at the start of the growing season, the observed pattern might also stem from seasonal recruitment patterns reflecting long-standing local coadaptation between early- and late-emerging genotypes. Recently, it has been shown that the microbiome can also structure $G \times G$ interactions, resulting in $G \times M \times G$ interactions and modified phenotypes with increased *Microcystis* tolerance (Macke et al., 2017). If, however, both *Daphnia* and *Microcystis* evolve in response to each other, the *Daphnia*–*Microcystis* interaction might be a nice example of a geographic mosaic of coevolution in which temporal adap-

tation result in spatial variation of coevolutionary hotspots and coldspots (Thompson, 2005). Modified locally adapted *Daphnia* microbiomes have been shown to mediate tolerance to *Microcystis* (Houwenhuyse et al., 2021). These aspects have high potential relevance with respect to the management of *Microcystis* blooms.

Overall, this study demonstrates the importance of short-term eco-(co)evolutionary dynamics in determining interactions between cyanobacteria and zooplankton grazers. We show that the composition of cyanobacteria and their toxicity for zooplankton grazers varies over short periods of time, and that the zooplankton grazer studied, the water flea *D. magna*, is able to adapt to these changes in order to reduce mortality imposed by *Microcystis*. This mechanism likely contributes to the ability of *Daphnia* populations to delay or even prevent cyanobacteria blooms in freshwater systems (Benndorf & Henning, 1989; de Bernardi & Giussani, 1990; Jeppesen et al., 1996; Matveev et al., 1994; Søndergaard et al., 2008; Vanni, 1984). It may also explain why the literature on the toxicity of *Microcystis* and the capacity of *Daphnia* to graze on and control *Microcystis* is so divergent in its conclusions, as the outcome may depend on who of the two interacting partners in the particular setting studied is at the moment of sampling winning in the coadaptation, and because most studies combine *Microcystis* and *Daphnia* strains that do not co-occur. The eco-evolutionary dynamics that may result from the dynamics reported in our study might impact the observed seasonality and food-web interactions in lakes and ponds (Schaffner et al., 2019).

In this study, the eco-evolutionary interactions between cyanobacteria and zooplankton grazers constitute a peculiar example of coevolution between antagonistic species (Thompson, 2005). Many examples of coevolution between antagonistic species have been documented, especially among host–parasite interactions (Anderson & May, 1982) and predator–prey interactions (Ehrlich & Raven, 1964), but also between competing species (Connell, 1980). The coevolution between the cyanobacterium *Microcystis* and its grazer, *Daphnia*, studied by us, involves a case where a prey species can kill its predator because of the production of toxins. It stands out as a case where both antagonistically interacting species can kill each other. We know many examples of toxic plants and predators that have evolved toxin tolerance (Ehrlich & Raven, 1964), but in most cases these dynamics are assumed to be relatively slow (Hiltunen & Becks, 2014). Here, we show that antagonistic species interactions between primary producer and its grazer can drive rapid coevolutionary dynamics on timescales of weeks, with a clear impact on ecosystem functioning.

Supplementary material

Supplementary material is available online at [Evolution Letters](#).

Data and code availability

The data and full code for the analysis is available through the following GitHub repository: https://github.com/MFajgenblat/Daphnia_Microcystis_temporal_dynamics.

Author contributions

A.B., L.D.M., and E.D. conceived the study. A.B., K.C., E.v.d.B. and I.V. performed the experiment. M.F. and M.C. lead the statistical analysis of the data. M.F., E.G., and L.D.M. lead the writing of the manuscript. R.A. provided critical feedback during the writing pro-

cess. All mentioned co-authors contributed substantially to the further refinements of the manuscript.

Funding

M.F. acknowledges funding by the Research Foundation Flanders through a FWO PhD FR fellowship (grant number 11E3222N). M.F. and L.D.M. acknowledge support from the PONDERFUL project, funded by the European Union's Horizon 2020 research and innovation program under grant agreement No. ID 869296. L.D.M. and E.D. acknowledge financial support from a KU Leuven research project (number C16/2023/003) and FWO G061824N.

Conflict of interest

The authors have no conflicts of interest to declare.

Acknowledgments

We gratefully acknowledge the NGO Vrienden van Heverlee-bos en Meerdaalwoud and its volunteers for providing access to Langerodevijver and granting permission for sampling, which was essential for this study. We thank Laura Herregodts and Cedric De Witte-Vroman for their valuable practical help in performing this experiment.

References

- Agrawal, M. K., Bagchi, D., & Bagchi, S. N. (2001). Acute inhibition of protease and suppression of growth in zooplankton, *moira macrocopa*, by microcystis blooms collected in Central India. *Hydrobiologia*, 464, 37–44. <https://doi.org/10.1023/A:1013946514556>
- Almeida, R. A., Fajgenblat, M., Lemmens, P., & De Meester, L. (2023). Pesticide exposure enhances dominance patterns in a zooplankton community. *Ecological Applications*, 33(7). e2900 <https://doi.org/10.1002/eap.2900>
- Anderson, R. M., & May, R. M. (1982). Coevolution of hosts and parasites. *Parasitology*, 85(2), 411–426. <https://doi.org/10.1017/S003182000055360>
- Asselman, J., De Coninck, D. I. M., Beert, E., Janssen, C. R., Orsini, L., Pfrender, M. E., Decaestecker, E., & De Schampelaere, K. A. C. (2017). Bisulfite sequencing with *Daphnia* highlights a role for epigenetics in regulating stress response to microcystis through preferential differential methylation of serine and threonine amino acids. *Environmental Science and Technology*, 51(2), 924–931. <https://doi.org/10.1021/acs.est.6b03870>
- Bates, D., Mächler, M., Bolker, B. M., & Walker, S. C. (2015). Fitting linear mixed-effects models using lme4. *Journal of Statistical Software*, 67(1), 1–48 <https://doi.org/10.18637/jss.v067.i01>
- Benndorf, J., & Henning, M. (1989). *Daphnia* and toxic blooms of *Microcystis aeruginosa* in Bautzen Reservoir (GDR). *Internationale Revue Der Gesamten Hydrobiologie Und Hydrographie*, 74, 233–248. <https://doi.org/10.1002/iroh.19890740302>
- Bogaerts, K., Komárek, A., & Lesaffre, E. (2017). *Survival analysis with interval-censored data. A practical approach with examples in R, SAS, and BUGS* (1st edition). Chapman and Hall/CRC. <https://doi.org/10.1201/9781315116945>
- Brans, K. I., & De Meester, L. (2018). City life on fast lanes: Urbanization induces an evolutionary shift towards a faster lifestyle in the water flea *Daphnia*. *Functional Ecology*, 32(9), 2225–2240. <https://doi.org/10.1111/1365-2435.13184>
- Capaul, M., & Ebert, D. (2003). Parasite-mediated selection in experimental *Daphnia magna* populations. *Evolution*, 57(2), 249–260. <https://doi.org/10.1111/j.0014-3820.2003.tb00260.x>
- Carpenter, B., Gelman, A., Hoffman, M., Lee, D., Goodrich, B., Betancourt, M., Brubaker, M., Guo, J., Li, P., & Riddell, A. (2017). Stan: A probabilistic programming language. *Journal of Statistical Software*, 76(1), 1–32. <https://doi.org/10.18637/jss.v076.i01>
- Carton, A. G. (2005). The impact of light intensity and algal-induced turbidity on first-feeding *Seriola lalandi* larvae. *Aquaculture Research*, 36(16), 1588–1594. <https://doi.org/10.1111/j.1365-2109.2005.01383.x>
- Chaturvedi, A., Zhou, J., Raeymaekers, J. A. M., Czapionka, T., Orsini, L., Jackson, C. E., Spanier, K. I., Shaw, J. R., Colbourne, J. K., & De Meester, L. (2021). Extensive standing genetic variation from a small number of founders enables rapid adaptation in *Daphnia*. *Nature Communications*, 12(1), 1–9. <https://doi.org/10.1038/s41467-021-24581-z>
- Chislock, M. F., Kaul, R. R. B., Durham, K. A., Sarnelle, O., & Wilson, A. E. (2019). Eutrophication mediates rapid clonal evolution in *Daphnia pulicaria*. *Freshwater Biology*, 64(7), 1275–1283. <https://doi.org/10.1111/fwb.13303>
- Connell, J. H. (1980). Diversity and the coevolution of competitors, or the ghost of competition past. *Oikos*, 35(2), 131. <https://doi.org/10.2307/3544421>
- Cuenca-Cambronero, M., Pantel, J. H., Marshall, H., Nguyen, T. T. T., Tomero-Sanz, H., & Orsini, L. (2021). Evolutionary mechanisms underpinning fitness response to multiple stressors in *Daphnia*. *Evolutionary Applications*, 14(10), 2457–2469. <https://doi.org/10.1111/eva.13258>
- de Bernardi, R., & Giussani, G. (1990). Are blue-green algae a suitable food for zooplankton? An overview. *Hydrobiologia*, 200–201(1), 29–41. <https://doi.org/10.1007/BF02530326>
- De Coninck, D. I. M., Janssen, C. R., & De Schampelaere, K. A. C. (2013a). An investigation of the inter-clonal variation of the interactive effects of cadmium and *Microcystis aeruginosa* on the reproductive performance of *Daphnia magna*. *Aquatic Toxicology*, 140–141, 425–431. <https://doi.org/10.1016/j.aquatox.2013.07.005>
- De Coninck, D. I. M., Janssen, C. R., & De Schampelaere, K. A. C. (2013b). An investigation of the inter-clonal variation of the interactive effects of cadmium and *Microcystis aeruginosa* on the reproductive performance of *Daphnia magna*. *Aquatic Toxicology*, 140–141, 425–431. <https://doi.org/10.1016/j.aquatox.2013.07.005>
- Decaestecker, E., De Gerssem, H., Michalakis, Y., & Raeymaekers, J. A. M. (2013). Damped long-term host-parasite Red Queen coevolutionary dynamics: A reflection of dilution effects? *Ecology Letters*, 16(12), 1455–1462. <https://doi.org/10.1111/ele.12186>
- Decaestecker, E., Gaba, S., Raeymaekers, J. A. M., Stoks, R., Van Kerkhoven, L., Ebert, D., & De Meester, L. (2007). Host-parasite “Red Queen” dynamics archived in pond sediment. *Nature*, 450(7171), 870–873. <https://doi.org/10.1038/nature06291>
- Dejenie, T., Asmelash, T., Rousseaux, S., Gebregiorgis, T., Gebrekidan, A., Teferi, M., Nyssen, J., Deckers, J., Van Der Gucht, K., Vyverman, W., De Meester, L., & Declerck, S. A. J. (2009). Impact of the fish Garra on the ecology of reservoirs and the occurrence of microcystis blooms in semi-arid tropical highlands: an experimental assessment using enclosures. *Freshwater Biology*, 54(8), 1605–1615. <https://doi.org/10.1111/j.1365-2427.2009.02209.x>
- Ebert, D. (2008). Host-parasite coevolution: Insights from the *Daphnia*-parasite model system. *Current Opinion in Microbiology*, 11(3), 290–301. <https://doi.org/10.1016/j.mib.2008.05.012>
- Ehrlich, P. R., & Raven, P. H. (1964). Butterflies and plants: A study in coevolution. *Evolution*, 18(4), 586–608. <https://doi.org/10.1111/j.1558-5646.1964.tb01674.x>

- Gandon, S., Buckling, A., Decaestecker, E., & Day, T. (2008). Host-parasite coevolution and patterns of adaptation across time and space. *Journal of Evolutionary Biology*, 21(6), 1861–1866. <https://doi.org/10.1111/j.1420-9101.2008.01598.x>
- Geerts, A. N., Vanoverbeke, J., Vanschoenwinkel, B., Van Doorslaer, W., Feuchtmayr, H., Atkinson, D., Moss, B., Davidson, T. A., Sayer, C. D., & De Meester, L. (2015). Rapid evolution of thermal tolerance in the water flea *Daphnia*. *Nature Climate Change*, 5(7), 665–668. <https://doi.org/10.1038/nclimate2628>
- Gelman, A., Meng, X.-L., & Stern, H. (1996). Posterior predictive assessment of model fitness via realized discrepancies. *Statistica Sinica*, 6(4), 733–760.
- Ger, K. A., Urrutia-Cordero, P., Frost, P. C., Hansson, L.-A., Sarnelle, O., Wilson, A. E., & Lüring, M. (2016). The interaction between cyanobacteria and zooplankton in a more eutrophic world. *Harmful Algae*, 54, 128–144. <https://doi.org/10.1016/j.hal.2015.12.005>
- Ghadouani, A., Pinel-Alloul, B., Plath, K., Codd, G. A., & Lampert, W. (2004). Effects of *Microcystis aeruginosa* and purified microcystin-LR on the feeding behavior of *Daphnia pulicaria*. *Limnology and Oceanography*, 49(3), 666–679. <https://doi.org/https://doi.org/10.4319/lo.2004.49.3.0666>
- Gianuca, A. T., Pantel, J. H., & De Meester, L. (2016). Disentangling the effect of body size and phylogenetic distances on zooplankton top-down control of algae. *Proceedings of the Royal Society B: Biological Sciences*, 283(1828), 20160487. <https://doi.org/10.1098/rspb.2016.0487>
- Gray, S. M., Bieber, F. M. E., McDonnell, L. H., Chapman, L. J., & Mandrak, N. E. (2014). Experimental evidence for species-specific response to turbidity in imperilled fishes. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 24(4), 546–560. <https://doi.org/10.1002/aqc.2436>
- Guillard, R. R. L., & Lorenzen, C. J. (1972). Yellow-green algae with chlorophyllide c. *Journal of Phycology*, 8(1), 10–14. <https://doi.org/10.1111/j.1529-8817.1972.tb03995.x>
- Gulati, R. D., Pires, L. M. D., & Van Donk, E. (2008). Lake restoration studies: failures, bottlenecks and prospects of new ecotechnological measures. *Limnologica*, 38(3–4), 233–247. <https://doi.org/10.1016/j.limno.2008.05.008>
- Gustafsson, S., Rengefors, K., & Hansson, L.-A. (2005). Increased consumer fitness following transfer of toxin tolerance to offspring via maternal effects. *Ecology*, 86(10), 2561–2567. <https://doi.org/10.1890/04-1710>
- Hairston, N. G., Holtmeier, C. L., Lampert, W., Weider, L. J., Post, D. M., Fischer, J. M., Cáceres, C. E., Fox, J. A., & Gaedke, U. (2001). Natural selection for grazer resistance to toxic cyanobacteria: Evolution of phenotypic plasticity? *Evolution*, 55(11), 2203–2214. <https://doi.org/10.1111/j.0014-3820.2001.tb00736.x>
- Hairston, N. G., Lampert, W., Cáceres, C. E., Holtmeier, C. L., Weider, L. J., Gaedke, U., Fischer, J. M., Fox, J. A., & Post, D. M. (1999). Rapid evolution revealed by dormant eggs. *Nature*, 401(6752), 446–446. <https://doi.org/10.1038/46731>
- Hansson, L. A., Gustafsson, S., Rengefors, K., & Bomark, L. (2007). Cyanobacterial chemical warfare affects zooplankton community composition. *Freshwater Biology*, 52(7), 1290–1301. <https://doi.org/10.1111/j.1365-2427.2007.01765.x>
- Hegg, A., Radersma, R., & Uller, T. (2022). Seasonal variation in the response to a toxin-producing cyanobacteria in *Daphnia*. *Freshwater Biology*, 67(6), 1035–1044. <https://doi.org/10.1111/fwb.13899>
- Hiltunen, T., & Becks, L. (2014). Consumer co-evolution as an important component of the eco-evolutionary feedback. *Nature Communications*, 5. <https://doi.org/10.1038/ncomms6226>
- Houwenhuysse, S., Stoks, R., Mukherjee, S., & Decaestecker, E. (2021). Locally adapted gut microbiomes mediate host stress tolerance. *ISME Journal*, 15(8), 2401–2414. <https://doi.org/10.1038/s41396-021-00940-y>
- Ilić, M., Cordellier, M., & Fink, P. (2021). Intrapopulation variability in a functional trait: Susceptibility of *Daphnia* to limitation by dietary fatty acids. *Freshwater Biology*, 66(1), 130–141. <https://doi.org/10.1111/fwb.13623>
- Isanta-Navarro, J., Hairston, N. G., Beninde, J., Meyer, A., Straile, D., Möst, M., & Martin-Creuzburg, D. (2021). Reversed evolution of grazer resistance to cyanobacteria. *Nature Communications*, 12(1). <https://doi.org/10.1038/s41467-021-22226-9>
- Jeppesen, E., Søndergaard, M., Jensen, J. P., Mortensen, E., & Sortkjaer, O. (1996). Fish-induced changes in zooplankton grazing on phytoplankton and bacterioplankton: a long-term study in shallow hypertrophic Lake Sebygaard. *Journal of Plankton Research*, 18(9), 1605–1625. <https://doi.org/https://doi.org/10.1093/plankt/18.9.1605>
- Jiang, X., Gao, H., Zhang, L., Liang, H., & Zhu, X. (2016). Rapid evolution of tolerance to toxic microcystin in two cladoceran grazers. *Scientific Reports*, 6(1), 25319. <https://doi.org/10.1038/srep25319>
- Kay, M. (2022). tidybayes: Tidy Data and Geoms for Bayesian Models (3.0.2). Github. <http://mjskay.github.io/tidybayes/>. Accessed: 02/01/2025
- Kosten, S., Huszar, V. L. M., Bécares, E., Costa, L. S., van Donk, E., Hansson, L. A., Jeppesen, E., Kruk, C., Lacerot, G., Mazzeo, N., De Meester, L., Moss, B., Lüring, M., Nöges, T., Romo, S., & Scheffer, M. (2012). Warmer climates boost cyanobacterial dominance in shallow lakes. *Global Change Biology*, 18(1), 118–126. <https://doi.org/10.1111/j.1365-2486.2011.02488.x>
- Lambrechts, L., Fellous, S., & Koella, J. C. (2006). Coevolutionary interactions between host and parasite genotypes. *Trends in Parasitology*, 22(1), 12–16. <https://doi.org/10.1016/j.pt.2005.11.008>
- Lampert, W., & Sommer, U. (2007). *Limnology: the ecology of lakes and streams* (2nd ed.). Oxford University Press. <https://doi.org/https://doi.org/10.1093/plankt/fbn013>
- Lemaire, V., Brusciotti, S., van Gremberghe, I., Vyverman, W., Vanoverbeke, J., & De Meester, L. (2011). Genotype × genotype interactions between the toxic cyanobacterium *Microcystis* and its grazer, the waterflea *Daphnia*. *Evolutionary Applications*, 5(2), 168–182. <https://doi.org/10.1111/j.1752-4571.2011.00225.x>
- Lemaire, V., Brusciotti, S., van Gremberghe, I., Vyverman, W., Vanoverbeke, J., & De Meester, L. (2012). Genotype × genotype interactions between the toxic cyanobacterium *Microcystis* and its grazer, the waterflea *Daphnia*. *Evolutionary Applications*, 5(2), 168–182. <https://doi.org/10.1111/j.1752-4571.2011.00225.x>
- Ley, R. E., Hamady, M., Lozupone, C., Turnbaugh, P. J., Ramey, R. R., Bircher, S. J., Schlegel, M. L., Tucker, T. A., Schrenzel, M. D., Knight, R., & Gordon, J. I. (2008). Evolution of mammals and their gut microbes. *Science*, 320(5883), 1647–1651. <https://doi.org/10.1126/science.1155390>
- Lüring, M. (2003). *Daphnia* growth on microcystin-producing and microcystin-free *Microcystis aeruginosa* in different mixtures with the green alga *Scenedesmus obliquus*. *Limnology and Oceanography*, 48(6), 2214–2220. <https://doi.org/10.4319/lo.2003.48.6.2214>
- Macke, E., Callens, M., De Meester, L., & Decaestecker, E. (2017). Host-genotype dependent gut microbiota drives zooplankton tolerance to toxic cyanobacteria. *Nature Communications*, 8(1). <https://doi.org/10.1038/s41467-017-01714-x>
- Matveev, V., Matveeva, L., & Jones, G. J. (1994). Study of the ability of *Daphnia carinata* King to control phytoplankton and resist cyanobacterial toxicity: Implications for biomanipulation in Australia. *Marine and Freshwater Research*, 45(5), 889–904. <https://doi.org/10.1071/MF9940889>

- Meerhoff, M., Audet, J., Davidson, T. A., De Meester, L., Hilt, S., Kosten, S., Liu, Z., Mazzeo, N., Paerl, H., Scheffer, M., & Jeppesen, E. (2022). Feedback between climate change and eutrophication: Revisiting the allied attack concept and how to strike back. *Inland Waters*, 12(2), 187–204. <https://doi.org/10.1080/20442041.2022.2029317>
- Miner, B. E., De Meester, L., Pfrender, M. E., Lampert, W., & Hairston, N. G. (2012). Linking genes to communities and ecosystems: *Daphnia* as an ecogenomic model. *Proceedings of the Royal Society B: Biological Sciences*, 279(1735), 1873–1882. <https://doi.org/10.1098/rspb.2011.2404>
- Miner, Brooks E., & Kerr, Benjamin. (2010). Adaptation to local ultraviolet radiation conditions among neighbouring *Daphnia* populations. *Proceedings of the Royal Society B: Biological Sciences*, 278 (1710), 1306–1313. <https://doi.org/10.1098/rspb.2010.1663>
- Paerl, H. W., & Huisman, J. (2008). Blooms like it hot. *Science*, 320(5872), 56–57. <https://doi.org/10.1126/science.1156721>
- Paerl, H. W., & Otten, T. G. (2013). Harmful cyanobacterial blooms: Causes, consequences, and controls. *Microbial Ecology*, 65(4), 995–1010. <https://doi.org/10.1007/s00248-012-0159-y>
- Paerl, H. W., Hall, N. S., & Calandrino, E. S. (2011). Controlling harmful cyanobacterial blooms in a world experiencing anthropogenic and climatic-induced change. *Science of the Total Environment*, 409(10), 1739–1745. <https://doi.org/10.1016/j.scitotenv.2011.02.001>
- Pimentel, D., McNair, S., Janecka, J., Wightman, J., Simmonds, C., O'connell, C., Wong, E., Russel, L., Zern, J., Aquino, T., & Tsomondo, T. (2001). Economic and environmental threats of alien plant, animal, and microbe invasions. *Ecosystems and Environment*, 84, 1–20. [https://doi.org/https://doi.org/10.1016/S0167-8809\(00\)00178-X](https://doi.org/https://doi.org/10.1016/S0167-8809(00)00178-X)
- R Core Team. (2023). R: a language and environment for statistical computing (4.3.1). R Foundation for Statistical Computing. <https://www.R-project.org/>. Accessed: 02/01/2025
- Sarnelle, O., & Wilson, A. E. (2005). Local adaptation of *Daphnia pulex* to toxic cyanobacteria. *Limnology and Oceanography*, 50(5), 1565–1570. <https://doi.org/10.4319/lo.2005.50.5.1565>
- Schaffner, L. R., Govaert, L., De Meester, L., Ellner, S. P., Fairchild, E., Miner, B. E., Rudstam, L. G., Spaak, P., & Hairston, N. G. (2019). Consumer-resource dynamics is an eco-evolutionary process in a natural plankton community. *Nature Ecology and Evolution*, 3(9), 1351–1358. <https://doi.org/10.1038/s41559-019-0960-9>
- Scheffer, M., & Van Nes, E. H. (2007). Shallow lakes theory revisited: Various alternative regimes driven by climate, nutrients, depth and lake size. *Hydrobiologia*, 584(1), 455–466. <https://doi.org/10.1007/s10750-007-0616-7>
- Scheffer, M., Carpenter, S., Foley, J. A., Folke, C., & Walker, B. (2001). Catastrophic shifts in ecosystems. *Nature*, 413(6856), 591–596. <https://doi.org/10.1038/35098000>
- Schwarzenberger, A., D'Hondt, S., Vyverman, W., & von Elert, E. (2013). Seasonal succession of cyanobacterial protease inhibitors and *Daphnia magna* genotypes in a eutrophic Swedish lake. *Aquatic Sciences*, 75(3), 433–445. <https://doi.org/10.1007/s00027-013-0290-y>
- Schwarzenberger, A., Hasselmann, M., & Von Elert, E. (2020). Positive selection of digestive proteases in *Daphnia*: A mechanism for local adaptation to cyanobacterial protease inhibitors. *Molecular Ecology*, 29(5), 912–919. <https://doi.org/10.1111/mec.15375>
- Sommer, U., & Maciej Gliwicz, Z. (1986). The PEG-model of seasonal succession of planktonic events in fresh waters. *Archiv Für Hydrobiologie*, 106(4), 433–471. <https://doi.org/https://doi.org/10.1127/archiv-hydrobiol/106/1986/433>
- Sommer, U., Adrian, R., De Senerpont Domis, L., Elser, J. J., Gaedke, U., Ibelings, B., Jeppesen, E., Lürling, M., Molinero, J. C., Mooij, W. M., Van Donk, E., & Winder, M. (2012). Beyond the plankton ecology group (PEG) model: mechanisms driving plankton succession. *Annual Review of Ecology, Evolution, and Systematics*, 43, 429–448. <https://doi.org/10.1146/annurev-ecolsys-110411-160251>
- Søndergaard, M., Liboriussen, L., Pedersen, A. R., & Jeppesen, E. (2008). Lake restoration by fish removal: short- and long-term effects in 36 Danish lakes. *Ecosystems*, 11(8), 1291–1305. <https://doi.org/10.1007/s10021-008-9193-5>
- Steiner, C. F. (2004). *Daphnia* dominance and zooplankton community structure in fishless ponds. *Journal of Plankton Research*, 26(7), 799–810. <https://doi.org/10.1093/plankt/fbh067>
- Stoks, R., Debecker, S., Dinh Van, K., & Janssens, L. (2015). Integrating ecology and evolution in aquatic toxicology: Insights from damselfish. *Freshwater Science*, 34(3), 1032–1039. <https://doi.org/10.1086/682571>
- Thompson, J. N. (2005). Coevolution: the geographic mosaic of co-evolutionary arms races. *Current Biology*, 15(24), R992–R994. <https://doi.org/10.1016/j.cub.2005.11.046>
- Trubetskova, I. L., & Haney, J. F. (2006). Effects of differing concentrations of microcystin-producing *Microcystis aeruginosa* on growth, reproduction, survivorship and offspring of *Daphnia magna*. *Archiv Fur Hydrobiologie*, 167(1–4), 533–546. <https://doi.org/10.1127/0003-9136/2006/0167-0533>
- Vanni, M. J. (1984). Biological control of nuisance algae by *Daphnia pulex*: Experimental studies. *Lake and Reservoir Management*, 1(1), 151–156. <https://doi.org/10.1080/07438148409354502>
- Vehtari, A., Gelman, A., Simpson, D., Carpenter, B., & Burkner, P. C. (2021). Rank-normalization, folding, and localization: An improved R-hat for assessing convergence of MCMC. *Bayesian Analysis*, 16(2), 667–718. <https://doi.org/10.1214/20-BA1221>
- von Elert, E., Martin-Creuzburg, D., & Le Coz, J. R. (2003). Absence of sterols constrains carbon transfer between cyanobacteria and a freshwater herbivore (*Daphnia galeata*). *Proceedings of the Royal Society B: Biological Sciences*, 270(1520), 1209–1214. <https://doi.org/10.1098/rspb.2003.2357>
- Xu, J., Chang, X., MacIsaac, H. J., Zhou, Y., Li, J., Wang, T., Zhang, J., Wen, J., Xu, D., Zhang, H., & Xu, R. (2023). Is a lower-toxicity strain of *Microcystis aeruginosa* really less toxic? *Aquatic Toxicology*, 263, 106705. <https://doi.org/10.1016/j.aquatox.2023.106705>